

1 **Title:**

2 Fecal indicator and pathogenic bacteria and their antibiotic resistance in alluvial groundwater of
3 an irrigated agricultural region with dairies.

4

5 **Short title:**

6 Dairy groundwater microbiological quality

7

8 Xunde Li ^{1,2}, Edward R. Atwill ^{1,2}, Elizabeth Antaki ², Olin Applegate ³, Brian Bergamaschi ⁴,
9 Ronald F. Bond ^{1,2}, Jennifer Chase ^{1,2}, Katherine R. Ransom ³, William Samuels ⁵, Naoko
10 Watanabe⁶, Thomas Harter ^{3*}

11

12 ¹ Department of Population Health and Reproduction, ² Western Institute for Food Safety and
13 Security, ³ Department of Land, Air and Water Resources, University of California, One Shields
14 Avenue, Davis, CA, 95616, ⁴ U.S. Geological Survey, Sacramento, CA, ⁵ California Department
15 of Water Resources, Sacramento, CA, ⁶ Faculty of Engineering, Division of Energy and
16 Environmental Systems, Hokkaido University, Kita 13 Nishi 8, Kita-ku, Sapporo-shi, Hokkaido,
17 060-8628 Japan.

18

19 **Abbreviations:**

20 CAFOs: Concentrated animal feeding operations

21 CFU: Colony-forming unit

22 MPN: Most probable number

23

24 *** Corresponding author**

25 Thomas Harter, Ph.D.

26 Department of Land Air Water Resources

27 University of California, Davis

28 125 Veihmeyer Hall

29 Davis, CA 95616

30 Phone: (001) 530-752-2709

31 Email: thharter@ucdavis.edu

32 **Abstract**

33 Surveys of microbiological groundwater quality were conducted in a region with intensive
34 animal agriculture in California, USA. The survey included monitoring and domestic wells in
35 eight confined animal feeding operations (CAFOs), and 200 small (domestic and community
36 supply district) supply wells across the entire region. *Campylobacter* was not detected in
37 groundwater while *E. coli* O157:H7 and *Salmonella* were each detected in 2 of 190 CAFO
38 monitoring well samples. Non-pathogenic generic *E. coli* and *Enterococcus* spp. were detected in
39 24.2% (46/190) and 97.4% (185/190) groundwater samples from CAFO monitoring wells and
40 4.2% (1/24) and 87.5% (21/24) of CAFO domestic wells, respectively. Concentrations of both
41 generic *E. coli* and *Enterococcus* spp. were significantly associated to well depth, seasons, and
42 the type of adjacent land use in the CAFO. No pathogenic bacteria were detected in groundwater
43 from 200 small supply wells in the extended survey. However, 4.5% to 10.3% groundwater
44 samples were positive for generic *E. coli* and *Enterococcus*. Concentrations of generic *E. coli*
45 were not significantly associated to any factors but concentrations of *Enterococcus* were
46 significantly associated to proximity to CAFOs, seasons, and concentrations of potassium in
47 water. Among a subset of *E. coli* and *Enterococcus* isolates from both surveys, the majority of *E.*
48 *coli* (63.6%) and *Enterococcus* (86.1%) isolates exhibited resistance to multiple (≥ 3) antibiotics.
49 Findings confirm significant microbial and antibiotic resistance loading to CAFO groundwater.
50 Results also demonstrate significant attenuative capacity of the unconfined alluvial aquifer
51 system with respect to microbial transport.

52

53 **Keywords:** CAFOs, dairy, groundwater, bacteria, antibiotic resistance

54

55 **Introduction**

56 Fecal-rich environments in concentrated animal feeding operations (CAFOs) are the pools
57 and potential sources of a wide variety of zoonotic pathogens (Hoar et al., 1999; Purdy et al.,
58 2001; Duffy, 2003; Lewis et al., 2005; Friesema et al., 2011; Won et al., 2013). For example,
59 dairy cattle provide natural reservoirs of *Campylobacter* (Dodson and LeJeune, 2005), *E. coli*
60 O15:H7 (Shere et al., 1998; Dodson and LeJeune, 2005), *Salmonella* (Dodson and LeJeune,
61 2005; Cummings et al., 2010), and *Cryptosporidium* (Atwill et al., 1998). Feces with high
62 concentrations of microbes are widely dispersed throughout CAFO environments, including
63 flush lane, corrals, pens, excise fields, floors and solid and liquid manure storage areas, etc.
64 (Lewis et al., 2005; Beck et al., 2007; Edrington et al., 2009; Toth et al., 2011; Watson et al.,
65 2012). Microbes from fecal-rich environments may reach groundwater via multiple routes
66 including, but not limited to surface runoff entrainment of feces deposited on the ground, leaking
67 of solid and liquid manure storage or storage areas, and subsurface transport (Harter et al., 2014).
68 CAFOs are of increasing concern for their impact on public health and the environment
69 including microbiological quality of groundwater (Kirkhorn, 2002; Bartelt-Hunt et al., 2011;
70 Lockhart et al., 2013). Coliform bacteria are known to be widely distributed in ground-water
71 (Embrey and Runkle, 2006). *E. coli* is able to travel long distances underground and is a useful
72 indicator of fecal contamination of groundwater (Foppen and Schijven, 2006).

73 Furthermore, the prevalence of antibiotic resistant bacteria has been well documented in
74 dairy animals (Fessler et al., 2012; Lindeman et al., 2013; Saini et al., 2013; Cummings et al.,
75 2014; Dues et al., 2014; Gibbons et al., 2014; Wichmann et al., 2014). The occurrence of
76 antibiotic resistant bacteria in animal production systems raises the potential for promoting
77 multiple-drug resistant bacteria (Esiobu et al., 2002; Straley et al., 2006; Wilhelm et al., 2009;

78 Holmes and Zadoks, 2011) and the transmission of antibiotic resistant bacteria to humans
79 through agriculture, the food chain, and the environment (Witte, 1998; Kummer, 2003; Ward et
80 al., 2014; Wieler, 2014).

81 Monitoring of fecal indicator and pathogenic bacteria in groundwater is important for
82 assessing the risk of microbial contamination of groundwater, especially in regions potentially
83 influenced by CAFOs. We first conducted a pilot survey to estimate the loads of fecal indicator
84 bacteria and pathogenic bacteria in the environment of two CAFOs. We then conducted a
85 systematic survey of indicator bacteria (generic *E. coli* and *Enterococcus*), pathogenic bacteria
86 (*Campylobacter*, *E. coli* O157:H7, and *Salmonella*), and antibiotic resistance in groundwater.
87 The survey was conducted at four groundwater transport scales: a) in groundwater immediately
88 below the water table at the dairy sites; b) in production aquifers immediately below dairies, c) in
89 production aquifers within the vicinity of dairies, and d) in production aquifers away from
90 dairies.

91 The study was conducted in the unconfined alluvial aquifer system of the Central Valley of
92 California, USA, which underlies an irrigated agricultural region with a large number of dairy
93 CAFOs (Figure 1). The survey included repeated, seasonal sampling events in monitoring and
94 domestic wells of eight commercial dairies followed by a broader survey of private domestic
95 wells across the region. A subset of generic *E. coli* (gram negative) isolates and *Enterococcus*
96 (gram positive) isolates from groundwater collected in these surveys were assessed for their
97 susceptibility to antibiotics. The objective of our work was to determine the frequency and
98 magnitude of indicator and pathogenic bacteria and their antimicrobial resistance in groundwater
99 at various distances from their source, assess risk factors related to microbial contamination of
100 groundwater, and further determine antibiotic resistance characteristics of bacteria in

101 groundwater. The working hypothesis was that wells with close proximity to CAFOs are more
102 vulnerable to microbial contamination and antibiotic resistance.

103

104 **Materials and Methods**

105 **Study area**

106 The Central Valley is an area of intensive agricultural production with 3 million ha, nearly
107 two-thirds of the total land area, devoted to irrigated farming (Burow et al., 2008). Sources of
108 irrigation water include both groundwater and surface water (Faunt et al., 2009). Irrigated crops
109 on or near dairies include leafy greens used for human consumption. Rural communities and
110 households and many urban areas rely on groundwater as their sole source of drinking water,
111 with minimal or no water treatment. Microbial contamination of groundwater is therefore a
112 significant concern for food safety and human health in this region.

113 The study area was comprised of the four counties with the largest concentration of dairies in
114 the California Central Valley: Stanislaus, Merced, Tulare, and Kings Counties (Figure 1). The
115 underlying Central Valley aquifer system is formed by unconsolidated alluvial fan and fluvial
116 basin sediments of varying quaternary and late tertiary ages. These sediments comprise the upper
117 500 to 1,000 m of thicker continental and underlying, older marine sediments (DWR, 2004; Page
118 et al., 1986). Hydraulic conductivity can vary greatly depending on the particle size of
119 sediments: coarse fraction hydraulic conductivity and fine fraction hydraulic conductivity have
120 been estimated to be $1,000 \text{ m d}^{-1}$ and less than 0.1 m d^{-1} , respectively (Faunt et al., 2009). The
121 Central Valley is broadly divided into three contiguous sub-basins, the northern Sacramento
122 Valley, the southcentral San Joaquin Valley (SJV) and the southern Tulare Lake Basin (TLB)

123 (Gronberg et al., 1998). Stanislaus and Merced Counties are within the SJV, while Tulare and
124 Kings Counties are within the TLB.

125 Depth to the water table varies and is thought to have significant impact on microbial
126 transport. Depth to water table near the Sierra foothills in Stanislaus and Merced Counties, in
127 spring 2010, was between 50-80 m below ground surface (bgs) and decreased in a southwesterly
128 direction to between 3-15-m bgs near the valley axis (Thalweg) formed by the San Joaquin River
129 (DWR, 2012). Depth to unconfined and semi-confined groundwater in Tulare and Kings
130 Counties, in spring 2010, generally increased from 10-15 m bgs in northeastern Tulare County to
131 over 100 m bgs in southern Tulare County and to 50-80 m bgs in Kings County and eastern
132 Tulare County (DWR, 2012).

133 The Central Valley has a Mediterranean climate with hot, dry summers and a rainy season
134 typically lasting from November through April. Average annual precipitation in the study area is
135 310 mm. The region supports 250 including tree fruit, nuts, vineyards, vegetables, rice, cotton,
136 and forage crops (corn, sorghum, grains, and alfalfa). Approximately 1.7 million mature cows
137 plus support cattle, about three-quarters of the California dairy herd, are located on less than
138 1,500 dairy farms (United States Environmental Protection Agency, 2012), mostly in the SJV
139 and TLB portion of the Central Valley.

140 Dairies are operated as CAFOs; they house mature animals in freestalls with exercise yards
141 (freestall dairies) or in open lots (drylot dairies). Dairies collect stormwater runoff from their
142 corrals (exercise yards, open lots, other animal holding areas) and washwater from their milking
143 barn in storage lagoons. Animal waste (manure) in freestall dairies is collected in concrete lanes
144 that are frequently flushed with recycled storage lagoon water. Manure solids are mechanically
145 separated from flushwater (freestall dairies) and scraped from corrals (freestall and drylot

146 dairies). Manure solids are dried, stored, and used for animal bedding or on cropland.
147 Wastewater from storage lagoons is typically applied to cropland via pipes and mixed with flood
148 irrigation water. Dairies typically manage a significant amount of forage crop acreage, which is
149 where most manure is applied. All management units of a dairy are subject to some leaching and
150 groundwater recharge – forage fields treated with manure, lagoons, and corrals (Harter et al.,
151 2002; Van der Schans et al., 2009). “Corrals” here include unlined freestalls, drylots, exercise
152 yards, hospital barns, and calf and heifer housing areas.

153

154 **Fecal indicator and pathogenic bacteria occurrence at land surfaces in dairies**

155 Between 2006 and 2008, five sampling events were conducted in two commercial dairies in
156 the Central Valley, California, for source characterization. During each sampling event, surface
157 solid and flush water (wastewater) samples were collected from each of several management
158 units in each dairy; solid samples were taken from manure fields, calf hutches, lactating cow
159 freestalls, lactating cow exercise yards, hospital pens, and heifer yards. Flushwater samples were
160 taken from storage lagoons, flush alleys in lactating cow freestalls, and a flush alley draining the
161 calf hutch area. Water samples were collected by directly pouring into sterilized 1-L
162 polyethylene bottles, while solid samples were collected using sterilized forceps and placed into
163 sterilized 1-L polyethylene bottles. Within each management unit, 12 randomly distributed
164 samples were collected, combined and thoroughly mixed for the final composite sample at each
165 sampling date. All Samples were kept cool in an ice chest while in the field and during
166 transportation to the laboratory, stored in a cold room (4°C) upon arrival at the laboratory, and
167 processed within 24-h after collection. One gram of solid samples or 1.0-ml of water samples
168 were suspended in PBS in 50-ml tubes and homogenized by shaking for 15-min using a wrist

169 action shaker. After shaking, solid particulates were removed by filtering through four-layer
170 gauze in a funnel and filtrates were 10-fold serially diluted. Dilutions were filtered using the
171 membrane filtration method for detection of generic *E. coli*, *Enterococcus*, and *Campylobacter*
172 as described below. For quantitative detection of *Salmonella*, ×4 replicates of each weight or
173 volume were suspended in 50-m L of buffered peptone water (BPW):10.0-g, 1.0-g, and 0.1-g of
174 solid samples or 1.0-ml, 0.1-ml, and 0.01-ml of water samples, followed by the MPN (most
175 probable number) method described below.

176

177 **On-Dairy groundwater monitoring**

178 Between 2008 and 2009, eight commercial dairy farms were enrolled for the groundwater
179 monitoring survey based on voluntary participation. Two dairies were located in Stanislaus
180 County in a region with highly permeable loamy sand and sandy loam soils and with a relatively
181 shallow water table (about 3-m bgs). Two dairies are located in Kings County and four dairies in
182 Tulare County, all over clayey to sandy loam soils with depth to groundwater ranging from 15 m
183 to over 30 m. On each dairy farm, groundwater samples were collected from 5.1 cm diameter
184 PVC monitoring wells constructed with 3 m to 6 m long well screens in the first non-clayey
185 alluvial sediment unit below the water table. Monitoring wells were constructed immediately
186 downgradient from manure-treated fields, storage lagoons, and corrals.

187 On-dairy groundwater samples were also obtained from domestic wells, which are typically
188 constructed with screens that are 10 m or more below the water table (Lockhart et al., 2013).
189 Wells were sampled seasonally, once during the coldest part of the rainy season (January 2008),
190 twice at the end of the rainy season (April 2008, March-April 2009), and once toward the end of
191 the hot, dry season (September 2008). Not all monitoring wells were always accessible or

192 available of water. In total, 190 samples were collected from 46 monitoring wells and 24 samples
193 were collected from 5 domestic on-dairy wells.

194

195 **Near-Dairy and Non-Dairy groundwater monitoring**

196 In 2010 and 2011, we extended our survey to general private domestic wells including six
197 small community service district wells across the four county regions. Domestic wells were
198 chosen based on responses from property owners to newspaper ads and to flyers mailed to rural
199 residents. In total, 200 domestic wells were enrolled (half in the SJV and half in the TLB) and
200 each well was sampled once between summer 2010 and summer 2011 (Lockhart et al., 2013).
201 Spatial analysis was used to determine the distance between a well and the nearest dairy corral or
202 lagoon. Wells located within 2.4-km from a dairy corral or dairy storage lagoon, including 12
203 domestic wells located on previously unsampled dairy properties, were classified as “near-dairy
204 wells” (132 wells), otherwise, they were classified as “non-dairy” (68 wells). Non-dairy wells
205 were considered to have low likelihood to have dairy management units within their recharge
206 source area. All wells were located in the vicinity of irrigated agricultural lands, some of which
207 may have manure applied by growers for soil amelioration (Lockhart et al., 2013).

208

209 **Groundwater sampling and filtering**

210 We developed and tested a novel approach for collecting microbial groundwater quality
211 samples from dairies. Details of the microbial field sampling protocol for monitoring wells are
212 described in Harter et al. (2014). For monitoring wells a portable, submersible, variable speed,
213 stainless steel Grundfos™ RediFlo2 pump (Enviro-Equipment, Inc.) was used. Purging volumes
214 prior to sampling ranged from 13 to 18 well volumes (about 190 L). At domestic or small

215 community service district wells, samples were collected with a closed, air-tight sampling
216 system. Samples were collected from spigots before the storage tank when possible, or at the
217 closest accessible spigot to the wellhead. Purging volumes ranged from 60 – 400 L. Water
218 samples were kept cool in an ice chest while in the field and during transportation to the
219 laboratory, stored in a cold room (4°C) upon arrival at the laboratory, and processed within 24 h
220 after collection.

221 For water samples collected on-dairy between 2008 and 2009, the default volume of water
222 filtered for generic *E. coli*, *Enterococcus*, *E. coli* O157:H7 and *Campylobacter* was 10 L for
223 each microbial analyte (40 L total) with occasionally smaller volume filtered for turbid water
224 samples. Water was filtered using a 10 L dispensing pressure vessel system (EMD Millipore
225 Corporation, Billerica MA) through 142-mm diameter 0.45- μ m pore size nitrocellulose membrane
226 filters as previously described (Li et al., 2014). To ensure numbers of colonies on plates were
227 countable for samples with high concentrations, additional 100-ml was filtered for generic *E. coli*
228 and 100-ml and 1-ml were filtered for *Enterococcus* through 47-mm diameter 0.45- μ m pore size
229 nitrocellulose membrane filters using a membrane filtration method. For quantitative detection of
230 *Salmonella*, $\times 4$ replicates of each volume were filtered: 2000-ml, 200-ml, and 20-ml.

231 For water samples collected in near-dairy and non-dairy locations between 2010 and 2011, a
232 50 L water sample was collected and immediately concentrated using a hollow-fiber
233 ultrafiltration (UF) technique (also called tangential flow) that has been reported to be effective
234 for recovering a diverse array of microbes from water (Hill et al, 2005). The ultrafiltration was
235 conducted using single-use F200NR dialysis filters (Fresenius Medical Care, Lexington, MA);
236 samples were concentrated to ~1000-ml (retentate). Each retentate was split to 5% for
237 *Enterococcus*, 15% for generic *E. coli*, 25% for *Salmonella*, 25% for *Campylobacter* and 30%

238 for *E. coli* O157:H7. The retentates used for generic *E. coli* and *Enterococcus* were further split
239 into two aliquots of 5% and 95% respectively to ensure countable colonies on plates. All
240 retentates were filtered through 47-mm diameter 0.45- μ m pore size nitrocellulose membrane
241 filters. For quantitative detection of *Salmonella*, $\times 4$ replicates of each volume were filtered 50-
242 ml, 5-ml, and 0.5-ml.

243 For all groundwater samples, electrical conductivity, pH, temperature, and dissolved oxygen
244 were measured in the field using a YSI[®] 556Multi-Parameter Water Quality sensor. Separate
245 water samples were collected for laboratory analysis of nitrate plus nitrite and major dissolved
246 ions including potassium and sodium (APHA, 2005, US EPA 1993, 2015). Depth to water was
247 measured prior to sampling monitoring wells. Approximately 40% property owners provided
248 information of domestic well structure and water table depth (Lockhart et al., 2013).

249

250 **Detection of fecal indicator and pathogenic bacteria**

251 Immediately after filtration, filters were placed onto CHROMagar EC plates for detection of
252 generic *E. coli*, mEI *Enterococcus* Indoxyl- β -D-Glucoside agar plates for detection of
253 *Enterococcus*, Rainbow and MacConkey agar plates for detection of *E. coli* O157:H7, and
254 Campy-Line agar (CLA) for detection of *Campylobacter*. CHROMagar EC plates were
255 incubated at 35°C for 2 h followed by incubation at 44.5°C for 24 h; mEI plates were incubated
256 at 41.0°C for 24 ~ 48 h; Rainbow and MacConkey plates were incubated at 37°C for 24 h; and
257 CLA plates were incubated in an anaerobic chamber at 42.0°C for 48h. After incubation,
258 presumptive bacterial colonies were confirmed by biochemical tests and/or molecular analysis.
259 Generic *E. coli* was confirmed by biochemical tests including Indole, Triple Sugar Iron (TSI),
260 Urea, and Simmons Citrate, and Methyl Red–Voges-Proskauer (MR-VP); *Enterococcus* was

261 confirmed by biochemical tests including Brain Heart Infusion agar, Brain Heart Infusion Broth
262 (BHIB), BHIB with 6.5% NaCl, and Bile Esculin reactions. Confirmation of *Campylobacter* was
263 done by biochemical tests and gram stain morphology for dairy samples collected between 2006
264 and 2007, and by biochemical tests and molecular analysis for water samples collected in
265 subsequent years. The biochemical tests for *Campylobacter* included Hippuric Acid, Oxidase,
266 and Catalase reactions. For molecular analysis, we used a specific PCR described previously
267 (Fermer and Engvall, 1999) to identify thermophilic campylobacters. *E. coli* O157:H7 was
268 confirmed by PCR using primers and PCR conditions described by Paton and Paton (2003).
269 Concentrations of confirmed bacteria for each sample were calculated and expressed as number
270 of CFU/g or ml for dairy surface solid and water samples and as number of CFU/100-ml for
271 groundwater samples.

272 For enumeration of *Salmonella*, 142-mm and 47-mm filters were inserted into 20-ml or 5-ml
273 BPW respectively and incubated at 37°C for 24 h. Following incubation 10- μ L of BPW
274 enrichment was transferred to 1-ml of RV and incubated. Five μ L of the RV enrichment was
275 plated onto XLD agar. Presumptive *Salmonella* colonies were confirmed biochemically using
276 TSI, Urea, and Lysine Iron Agar. The numbers of confirmed positive reactions of each filtration
277 (volume and replicate) were used for calculating *Salmonella* concentrations using a MPN
278 calculator (Curiale) and expressed as MPN/g or ml for dairy surface solid and water samples and
279 as MPN/100-ml for each groundwater sample.

280

281 **Antibiotic resistance assay of indicator bacteria**

282 Antibiotic resistant profiles were determined for a subset of generic *E. coli* and *Enterococcus*
283 obtained from groundwater. A gram negative (G-) Sensititre® plate (CMV2AGNF) and a Gram

284 positive (G+) Sensititre® plate (CMV3AGPF) (Trek Diagnostic Systems Inc., Westlake, OH)
285 were used for *E. coli* and *Enterococcus* respectively, according the manufactures instructions. *E.*
286 *coli* strains (ATCC 25922, ATCC35218) and *Enterococcus* strain (ATCC29212) were used as
287 quality control strains. The Minimum Inhibitory Concentration (MIC) values were the lowest
288 concentrations of antibiotics that inhibit visible growth of bacteria. Interpretations of antibiotic
289 resistance were set by the criteria of the MIC breakpoints developed by the Veterinary
290 Antimicrobial Susceptibility Testing Subcommittee of the Clinical and Laboratory Standards
291 Institute (CLSI) (Watts, 2008). An isolate of bacteria is defined as multiple-drug resistant if the
292 isolate is resistant to ≥ 3 antibiotics.

293

294 **Statistical analysis**

295 Because *Campylobacter* was not detected and *Salmonella* and *E. coli* O157:H7 were each
296 only detected in two samples, statistical analyses were conducted on generic *E. coli* and
297 *Enterococcus* for on-dairy, near-dairy and non-dairy water samples. Mean concentrations of
298 generic *E. coli* and *Enterococcus* were calculated and evaluated using one-way ANOVA
299 (Minitab, Minitab Inc, State College, Pennsylvania) tests to determine statistical differences
300 between bacterial concentrations within well types and designations. Significance was set at
301 $P \leq 0.05$ for each test. For on-dairy samples, the association between bacteria concentrations and
302 well types (domestic vs. monitoring well), the primary dairy facility component in upgradient
303 proximity to wells, relative to groundwater flow (lagoon, corral, manure-treated field), water
304 table depth (shallow: less than 5-m; deep: 13-m to over 30-m) and season (winter - Jan'08,
305 spring - Apr'08, Mar'09, and fall - Sept'08) were analyzed using Poisson regression (STATA 12
306 software, College Station, Texas), with $P \leq 0.05$ for inclusion in the final model. For near-dairy

307 and non-dairy samples, the association between bacteria concentrations and well location (near-
308 dairy well vs. non-dairy well), the distance to the nearest dairy corral or lagoon (disregarding
309 groundwater gradients), field temperature, pH, salinity, solute concentrations, and sampling
310 event season were analyzed using the same Poisson Regression. Two wells were excluded as
311 outliers, showing extreme differences in bacterial counts relative to the overall dataset. We did
312 not consider groundwater flow direction in the determination of domestic well distance to dairy,
313 since gradients are highly variable, transient, and often controlled by seasonal irrigation wells
314 that pump at rates exceeding $15 \text{ m}^3 \text{ min}^{-1}$.

315

316 **Results**

317 **Survey of fecal indicator and pathogenic bacteria at land surfaces on dairies**

318 Surveys of fecal indicator and pathogenic bacteria concentrations on land surfaces in CAFO
319 dairies are shown in Figure 2. Figure 2 also illustrates two typical variations of the spatial layouts
320 of Central Valley freestall dairy management units, albeit without the complete layout of
321 manure-treated fields surrounding these dairies. The two fecal indicator bacteria, generic *E. coli*
322 and *Enterococcus* were widely distributed in all solid and water samples of the various surface
323 environments in the two CAFOs. Typical concentrations of these two bacterial indicators in the
324 above-ground matrices (liquid manure slurries and solids) ranged from hundreds of thousands to
325 over two million CFU/100-ml slurry or CFU/g solids. Despite their ubiquitous occurrence
326 through dairy management units in contact with manure, much lower concentrations of indicator
327 bacteria were found in the control fields next to each dairy that were not treated with manure (2
328 to >3 orders of magnitude less).

329 The primary environmental load of *Campylobacter* appeared to be liquid manure slurries and
330 not the large amount of surface manure solids present on the dairies: *Campylobacter* was
331 detected in slurries at concentrations typically between 10^2 and 10^4 CFU/100-ml while it was
332 detected in only a single sample of surface solids. *Salmonella* counts in liquid manure samples
333 were generally lower, compared to *Campylobacter* and also appeared to have high temporal
334 variability between sampling events. In contrast to *Campylobacter*, *Salmonella* was detected
335 more frequently, if only at low levels, in surface solids on the dairy, particularly in the shaded
336 hospital pen and freestall structures. With the exception of *Salmonella* in April '07, no pathogens
337 were detected at any time in control fields.

338

339 **On-Dairy monitoring of fecal indicator and pathogenic bacteria in groundwater**

340 In on-dairy groundwater samples, *Campylobacter* was neither detected in groundwater
341 immediately below the water table (monitoring wells), nor in domestic wells, which tap
342 groundwater at several tens to over one hundred meter below the water table. In contrast
343 *Salmonella* and *E. coli* O157:H7, while not present in domestic well water, each occurred in 1%
344 (2/190) of monitoring well samples.

345 The two *Salmonella* detections occurred during the winter sampling, in January 2008. One
346 monitoring well, with a low concentration of 0.04 MPN/100-ml, was located downgradient of a
347 typical manure-treated field with sandy loam soil and relatively shallow 5 m depth to water table
348 on a dairy located in the SJV. Nitrate and salinity show significant influence from manure
349 applications, but are not as high as in other wells located downgradient from manure-treated
350 fields on this or nearby dairies described in Harter et al. (2002). Hence, the well does not appear
351 exceptionally vulnerable to manure leaching. The other monitoring well, with a concentration of

352 0.02 MPN/100 ml, was located adjacent to a corral on a TLB dairy overlying 27 m of
353 unsaturated, highly heterogeneous, sandy, loamy, and clayey alluvial sediments. The monitoring
354 well is screened from the water table at 27 m to 35 m. Total nitrogen (7 mg L^{-1}) and salinity are
355 lower than at nearby wells and do not indicate strong manure influence, but may be influenced
356 by recharge from a nearby (150 m) unlined irrigation canal.

357 The two *E. coli* O157:H7 detections occurred during sampling in March 2009. One sample
358 came from the same well that was positive for *Salmonella* 14 months earlier. The second
359 detection was in a well located adjacent to a freestall flush lane, in a nearby SJV dairy. At both
360 locations, the water table is relatively shallow, at 3-5 m below ground surface.

361 Table 1 shows the survey results of generic *E. coli* in groundwater in CAFOs. Among the 24
362 samples collected over the 4 sampling events from on-dairy domestic wells, only 1 was positive
363 for generic *E. coli* with a concentration of 0.01 CFU/100 ml. This sample, from a dairy in the
364 SJV study area, was obtained from a well where depth to ground water varies (3-5 m) and which
365 had an unknown screened interval. In contrast, among on-dairy monitoring wells 24.2% (46/190)
366 of the water samples were positive for generic *E. coli* with a range of 15.2%-27.5% between
367 different seasons. Generic *E. coli* was not detected in monitoring wells at two relatively new
368 (<10 year old) dairy farms with depth to groundwater exceeding 20 m (however, one of these
369 was only sampled in January 2008). Depending on season and farm, mean concentrations of
370 generic *E. coli* in monitoring wells ranged from 0.01 CFU/100 ml to 35.01 CFU/100 ml.

371 *Enterococcus* was detected in 97.4% (185/190) of water samples from monitoring wells
372 (Table 2). Despite their ubiquitous presence, concentrations mostly did not exceed 100 CFU/100
373 ml. Some extremely high concentrations were detected in monitoring wells at the two SJV
374 dairies with the shallow (<10 m) water table in March-April of 2009. In on-dairy domestic wells,

375 87.5% (21/24) of water samples tested positive for *Enterococcus*, but with overall lower
376 concentrations than in monitoring wells (Table 2).

377 The concentrations of both generic *E. coli* and *Enterococcus* were significantly associated
378 with the type of dairy land use immediately upgradient of monitoring wells, with the depth to
379 water table, and with season. Well type (domestic vs. monitoring) and, thus, depth of well screen
380 below the water table (immediately below the water table vs. production level groundwater) was
381 also a statistically significant factor (Table 3). *E. coli* did not occur in domestic wells at sufficient
382 rates to be included in the statistical model. In order to assess the association between dairy
383 management units and the occurrence of indicator bacteria, the distribution of types of land use
384 with proximity to wells and the frequency of detection of generic *E. coli* and *Enterococcus* in
385 water from monitoring wells were compared (Figure 3). The highest frequencies of detection of
386 both generic *E. coli* and *Enterococcus* were associated with monitoring wells immediately
387 downgradient of manure-treated fields and corrals. Monitoring wells downgradient of lagoons
388 had lower concentrations than others, but were higher than those of the (deeper screened) on-
389 dairy domestic wells.

390

391 **Monitoring of fecal indicator and pathogenic bacteria in drinking water supply wells**

392 We detected no pathogenic bacteria in any water samples from the 200 domestic wells
393 sampled in the 2010-2011 campaign, regardless whether the domestic well was nearby or further
394 away from a dairy ('near-dairy' vs. 'non-dairy'). However, 4.5% and 7.5% of near-dairy wells
395 were positive for generic *E. coli* and *Enterococcus*. Similarly, 5.9% and 10.3% of non-dairy
396 wells were positive for generic *E. coli* and *Enterococcus*, respectively (Table 4). Concentrations
397 of generic *E. coli* were not significantly related to the distance from the nearest corral or lagoon,

398 water quality parameters or seasons (statistical data not shown). But *Enterococcus* results were
399 significantly different between near-dairy wells and non-dairy wells, between seasons, and were
400 negatively correlated to potassium concentration (Table 5). Microbial indicators were not
401 significantly associated with other dissolved solutes or water quality parameters in groundwater,
402 including total dissolved solids concentration.

403

404 **Antibiotic resistance assay of a subset of indicator bacteria**

405 Although only small subsets of bacteria were tested, all isolates of generic *E. coli* and
406 *Enterococcus* demonstrated resistance to at least one antibiotic. Moreover, the majority of
407 generic *E. coli* isolates (63.6%) and *Enterococcus* isolates (86.1%) exhibited multi-drug
408 resistance (resistant to three or more drugs), regardless of well type (monitoring vs. domestic
409 wells on CAFOs of the on-dairy survey) or distance from a dairy (near-dairy vs. non-dairy) of
410 the well from which samples were collected and used for isolating *E. coli* and *Enterococcus*
411 (Table 6). Among the near-dairy domestic well, one generic *E. coli* and three *Enterococcus*
412 isolates came from domestic wells on dairy facilities not studied in 2006-2009. Like others, these
413 isolates exhibited multi-resistant properties. We found that generic *E. coli* were most often
414 resistant to azithromycin, chloramphenicol, trimethoprim/sulfamethoxazole, and tetracycline and
415 *Enterococcus* were most often resistant to tigecycline, quinupristin/dalfopristin, linezolid,
416 chloramphenicol, erythromycin, iprofloxacin, and tetracycline.

417

418 **Discussion**

419 The high level of fecal indicator bacteria in CAFO surface samples are consistent with what
420 we would expect given the large fraction of fecal solids mixed in with these samples, exceeding

421 50% on a wet weight basis in many samples. High occurrence rates of *E. coli* and *Enterococcus*
422 have also been found on dairies in the northeastern U.S. dairies (Pradhan et al., 2009).
423 *Enterococcus* has been found in surface water and groundwater impacted by a concentrated
424 swine feeding operation in the Mid-Atlantic United States (Sapkota et al., 2007). Similarly,
425 pathogenic bacteria including *Campylobacter*, *E. coli* O157:H7 and *Salmonella* have been
426 commonly detected in dairy environment elsewhere but at significantly lower concentration than
427 indicator bacteria (Murinda et al., 2004; Toth et al., 2013; Ravva and Sarreal, 2014). In our
428 survey, the primary source of *Campylobacter* among the various dairy management units is
429 difficult to discern – solids samples did not yield significant information, while freestall and
430 lagoon water, which consistently yield significant *Campylobacter* may originate, with the
431 exception of manure-treated fields, from any of the dairy management units shown in Figure 2.

432 Calf hutch flush water originates from tap water. Hence, the consistent occurrence of both
433 *Campylobacter* and *Salmonella* indicates that calf hutches are at least one of the contributing
434 sources of pathogens. *Salmonella* was also most common in liquid slurries, but also occurred in
435 the surface solids that had little exposure to direct sunlight (hospital barn and freestall lots). The
436 lack of pathogens on other surface solids is consistent with earlier findings (Nicholson et al.,
437 2005) probably due to inactivation after exposure to ambient conditions including higher
438 temperatures (Hoar et al. 1999; Sinton et al., 2007; Mariarty et al., 2011). Survival of pathogens
439 in the dairy environment depends on numerous complex environmental factors (Toth et al., 2011;
440 Ravva and Sarreal, 2014), reflected here by the lack of strong seasonal signatures, despite the
441 high contrast between hot, dry summers and moist, cool winters. The lower frequencies and
442 concentrations of *Campylobacter* and *Salmonella* in liquid samples, when compared to fecal

443 indicator concentrations may largely be due to those being shed only by infected animals, which
444 may represent only a fraction of the herd.

445 Indicator bacteria and pathogens occurring on dairy CAFOs may be subject to transport into
446 the environment surrounding dairies through surface runoff to streams, and through incidental or
447 intentional infiltration into and transport through unsaturated porous medium to groundwater
448 (Joy et al., 1998; Searcy et al., 2005; Park et al., 2012; Unc and Goss, 2014). Unc et al. (2012)
449 found at least three orders of magnitude reduction in *Enterococcus* concentration across the 3 m
450 unsaturated zone profile on one of the two SJV dairies. Li et al (2014) estimated attenuation rates
451 for generic *E. coli* ranging from 3 to 7 orders of magnitude using 2006-2008 surface samples
452 reported here and a limited number of groundwater samples collected concurrently with surface
453 samples (not included in this study).

454 There are distinct differences in generic *E. coli* and in *Enterococcus* detection frequencies
455 between monitoring wells located immediately upgradient of dairies, which are comparable to
456 those in domestic wells, and detection frequencies in monitoring wells located within dairies
457 (Figure 3). Monitoring wells downgradient of corrals and manure-treated fields have much
458 higher detection frequencies than those downgradient of lagoons. The difference may be due to
459 more attenuation of microorganisms by the fine-grained sludge layer commonly found on the
460 bed of storage lagoons than in the fractured and mechanically impacted corral surface. Due to
461 mechanical preparation (ploughing etc.), fields provide a more open surface with significantly
462 higher infiltration rates than either corrals or lagoon beds, and thus less filtration of colloidal
463 microorganisms. Similarly, the already low risk of pathogenic contamination may actually be
464 lowest in the vicinity of storage lagoons relative to other dairy management units.
465 Coincidentally, the two *Salmonella* occurrences were not associated with lagoon leakage.

466 In agreement with previous reports, we find that microbial groundwater contamination
467 generally decreases with increased well depth (Goss et al., 1998; Pitkänen et al., 2011).
468 Monitoring wells are in closer proximity to animal production areas and waste storage facilities,
469 while domestic wells are screened at some depth below the water table. Concentrations of the
470 most commonly found bacteria in both types of wells, *Enterococcus*, is therefore not surprisingly
471 significantly less in domestic wells on dairies than in their monitoring wells (Fig. 1). On the
472 other hand, similar to the survey conducted in private wells used for drinking water in
473 northeastern Ohio (Won et al., 2013), no significant correlation was found between *E. coli*
474 concentrations and potential pollution factors in our domestic wells survey.

475 Concentrations of *Enterococcus* were significantly associated with potassium concentration
476 in groundwater. The electrochemical properties of soil can alter the transportation of bacteria
477 (Unc and Goss, 2004), which may explain the relationship of *Enterococcus* and the
478 concentrations of potassium in groundwater. Ionic strength (as indicated by TDS) was not a
479 significant factor. The association may also be explained by the fact that highest potassium
480 concentrations are often found in the anaerobic shallow ammonium plumes emanating from
481 older storage lagoons overlying shallow groundwater (Harter et al., 2002). As mentioned above,
482 the lagoon bed may be a significant filter of microbial contaminants, thus becoming a source of
483 relatively low indicator bacteria counts while also being a source of high potassium
484 concentrations.

485 The main reason for the low detection rate of pathogens in monitoring wells and their
486 absence in domestic wells appears to be the strong attenuation in the unsaturated zone combined
487 with physical limits of detection: Assuming there is no inhibition within the assay, as few as 1
488 CFU per volume filtered can be detected with membrane or pressure vessel filtration direct

489 plating methods. For the various Salmonella MPN methods used for this work, the detection
490 limit varies from 0.00013-140 MPN/ml or g. Given the 3 to 7 order of magnitude attenuation
491 estimated from highly prevalent indicator organisms at our dairy sites (Li et al., 2014), we would
492 expect pathogen concentrations to follow a similar trend in reduction. Hence, given the lower
493 starting concentrations in manure slurries compared to generic *E. coli* and *Enterococcus* (Figure
494 2), pathogen concentrations would be expected to be mostly below detection limits of these
495 water assays. This is confirmed by the fact that even the shallow-most groundwater samples
496 below dairies did not yield any *Campylobacter* occurrence. *E. coli* O157:H7 and *Salmonella* are
497 each detected in 2 of 51 on-dairy monitoring wells across 8 dairies, but in only 1 of 4 sampling
498 campaigns. Also consistent with attenuation rates estimated from indicator organisms, we
499 detected no pathogenic bacteria in the survey of on-dairy, near-dairy, or non-dairy domestic
500 wells. This suggest that 3 m to 30 m of unsaturated alluvial sediments with silty sand, loamy
501 sand, fine sand, and sandy loam or finer materials provide significant protection from pathogenic
502 transport to the water table. Assuming human water consumption on the dairy is limited to
503 domestic wells, these data suggest that the risk of human waterborne illness from consumption of
504 domestic well water is very low. Given that normal water consumption patterns of children and
505 adults can range from 1 to 3 liters per day, using water from monitoring wells as a source of
506 drinking water or using other on-dairy sources for municipal purposes may pose an unacceptable
507 risk of waterborne transmission if not treated. Hence, groundwater supplies for drinking water,
508 typically obtained tens to over 100 meters below the water table are also well protected in these
509 landscape settings.

510 The high pathogenic loading at the land surface of dairy CAFOs may pose a significant risk
511 to groundwater in other hydrogeologic and well settings: Horn (Horn et al., 2009) recognized

512 that poor well seal construction may be a significant risk factor groundwater contamination.
513 Compromised wells allow for rapid transport through the gravel filter of a domestic well.

514 Also, soils with significant macropores (e.g., fractured clay, till) or of much less thickness
515 than 3 m overlying more vulnerable sand and gravel aquifers or highly fractured rock aquifers
516 may be at significant risk near similarly managed CAFOs. We note that most of the surveyed
517 domestic wells are also in the vicinity of a private onsite wastewater treatment system (septic
518 system) that may serve as a source for enteric indicator bacteria and pathogens (Bremer et al,
519 2012). This and the use of manure as soil amendment may explain the occurrence of indicator
520 bacteria at significant distances from dairies.

521 Generic *E. coli* and *Enterococcus* are among the commonly used indicator organisms for
522 monitoring microbiological quality of water (Edberg et al., 1997). It is generally assumed that
523 indicator microbial pollution poses a significant risk of pathogen occurrence due to similar
524 transport mechanisms (Goss et al., 2002). In the investigated alluvial systems, it appears that the
525 significant difference in concentration of indicator vs. pathogenic bacteria at dairy surfaces lead
526 to significant occurrence of indicator organisms, while the actual risk of pathogenic bacteria
527 occurrence is very low. On the other hand, the absence of indicator organisms is not a guarantee
528 of clean water. In the two cases of pathogen detection immediately below the water table,
529 samples were negative for generic *E. coli* and had average *Enterococcus* concentrations (data not
530 shown). Other studies have also reported the lack of correlation between fecal indicator bacteria
531 and pathogens in groundwater (Ferguson et al., 2012).

532 High spatial attenuation rates of *E. coli* and other fecal indicator bacteria through sandy
533 aquifers have been found by other studies (Knappett et al., 2012; Pang, 2009). Although we did
534 not conduct microbial source tracking studies to determine the sources of generic *E. coli* and

535 *Enterococcus* spp. in groundwater, the resistance to multiple veterinary or medical drugs among
536 a subset of these bacteria points to human- or animal-waste derived sources of antibiotic resistant
537 organisms. Presence of antibiotic-resistant bacteria within groundwater from CAFO-specific
538 monitoring wells suggests these are animal-derived, and from fecal-rich environment within
539 CAFOs. Several studies have documented antibiotic resistant bacteria in groundwater under the
540 influence of concentrated swine operations (Chee-Sanford et al., 2001; Anderson and Sobsey,
541 2006; Mackie et al., 2006; Koike et al., 2007; Sapkota et al., 2007).

542 A previous study, which collected information about antibiotics use on two participating
543 CAFOs (Watanabe et al., 2010), also sampled surface solids for antibiotics within the same
544 2006-2008 campaign. They detected varied antibiotic residues such as tetracycline, lincomycin,
545 trimethoprim, sulfadimethoxine, and sulfamethazine. According to a survey conducted by the
546 USDA Animal and Plant Health Inspection Service (APHIS), sulfonamide and tetracycline are
547 among the most common antibiotics used in dairies in the U.S. (retrieved info sheet from
548 APHIS). In the present study, we found that genetic *E. coli* and *Enterococcus* isolates were
549 resistant to many of these commonly-deployed and used antibiotics of the US industry.

550 Additionally, we found the antibiotic resistance patterns of generic *E. coli* within groundwater
551 samples were consistent with or similar to generic *E. coli* within surface water samples from
552 dairy CAFOs (Gibson and Schwab, 2011a,b; Li et al., 2014). Our findings suggest and indicate
553 that there is significant potential risk of groundwater contamination with antibiotic resistant
554 bacteria derived from CAFOs, even if the subsurface environment is not suitable to transmit
555 pathogenic bacteria.

556 An earlier study documented and pointed out public health implications regarding multiple
557 antibiotic resistance gram negative bacteria in rural groundwater supplies used as drinking water

558 source (McKeon et al., 1995). It remains unclear, to which degree onsite wastewater treatment
559 systems contribute to antibiotic resistance found in groundwater samples of domestic wells,
560 especially in areas further than 2.4 km away from dairies. Manure amendments are commonly
561 used in irrigated agriculture throughout the region. This suggests an alternative source of
562 antibiotic resistant bacteria outside the direct zone of influence from dairy facilities. For future
563 work, we propose surveying antibiotic use across dairies and assessing antibiotic resistance
564 within both G+ and G- bacteria from dairy environments. This could include groundwater and
565 surface water with commonly-used, site specific antibiotics.

566

567 **Conclusion**

568 In groundwater immediately below the water table and in groundwater at production depth in
569 this irrigated agricultural region overlying an alluvial aquifer, we detected *E. coli* in 15% - 27%
570 and *Enterococcus* in 80%-100% of groundwater in dairy CAFOs. Both indicator bacteria were
571 detected at much lower rates ($\leq 10\%$) in groundwater at near-dairy and non-dairy domestic wells
572 of the same region. The prevalence of *Enterococcus* was significantly associated to the influence
573 of dairy operations. We did not detect pathogenic bacteria within domestic wells, on-dairy, near-
574 dairy, or in non-dairy areas through use of filtrate from 10 L water samples and enrichment;
575 however, most isolates of *E. coli* and *Enterococcus* from production depth groundwater exhibited
576 multi-drug antibiotic resistance. These findings outline the broad reach of antibiotic resistant
577 bacteria in groundwater of this region. Applying Good Agricultural Practices (GAPs) on CAFOs
578 and improving well maintenance practices such as well seals (Rudolph et al., 1996) are among
579 several possible measures to prevent bacteria at CAFO surfaces from entering groundwater.
580 From a public health perspective, continuous and effective groundwater monitoring is important

581 for protection from residual microbiological risks associated with groundwater. Further work is
582 needed to better understand the sources, occurrence and public health implications of antibiotic
583 resistance in enterically-derived and/or environmental bacteria within groundwater
584 environments.

585

586 **Acknowledgement**

587 Funding for our research projects was provided by the California State Water Resources
588 Control Board Contract Number 03-244-555-01, 04-184-555-0, and 11-168-150. We gratefully
589 acknowledge the support and collaboration of the eight dairy landowners and of the many
590 domestic well owners who allowed us to sample wells on their property. We also acknowledge
591 the project support from JohnFranco Saraceno and Dana Erickson at the USGS. Comments from
592 two anonymous reviewers greatly improved the manuscript.

593

594 **References**

- 595 Anderson, M.E, and M.D. Sobsey. 2006. Detection and occurrence of antimicrobially resistant *E.*
596 *coli* in groundwater on or near swine farms in eastern North Carolina. *Water Sci.*
597 *Technol.* 54: 211-218.
- 598 American Public Health Association, 2005. Standard Method 2320 Titration, Standard Methods
599 21st edition.
- 600 Atwill, E.R., J.A. Harp, T. Jones, P.W. Jardon, S. Checel, and M. Zylstra. 1998. Evaluation of
601 periparturient dairy cows and contact surfaces as a reservoir of *Cryptosporidium parvum*
602 for calfhood infection. *Am. J. Vet. Res.* 59: 1116-1121.

603 Bartelt-Hunt, S., D.D. Snow, T. Damon-Powell, and D. Miesbach. 2011. Occurrence of steroid
604 hormones and antibiotics in shallow groundwater impacted by livestock waste control
605 facilities. *J. Contam. Hydrol.* 123: 94-103.

606 Beck, J.P., A. Heutelbeck, and H. Dunkelberg. 2007. Volatile organic compounds in dwelling
607 houses and stables of dairy and cattle farms in Northern Germany. *Sci. Total. Environ.*
608 372: 440-454.

609 Bremer, J. and T. Harter. 2012. Domestic wells have high probability of pumping septic tank
610 leachate, *Hydrol. Earth Sys. Sci* 16:2453-2467, doi:10.5194/hess-16-2453-2012.

611 Burow, K.R., B.C. Jurgens, L.J. Kauffman, S. P. Phillips, B.A. Dalgish, and J.L. Shelton. 2008.
612 Simulations of ground-water flow and particle pathline analysis in the zone of
613 contribution of a public-supply well in Modesto, Eastern San Joaquin Valley, California.
614 Scientific Investigations Report 2008-5035. U.S. Geological Survey.

615 Chee-Sanford, J.C., R. I. Aminov, I. J. Krapac, N. Garrigues-Jeanjean, and R. I. Mackie. 2001.
616 Occurrence and diversity of tetracycline resistance genes in lagoons and groundwater
617 underlying two swine production facilities. *Appl. Environ. Microbiol.* 67: 1494-1502.

618 Cummings, K.J., V.A. Aprea, and C. Altier. 2014. Antimicrobial resistance trends among
619 *Escherichia coli* isolates obtained from dairy cattle in the northeastern United States,
620 2004-2011. *Foodborne Pathog. Dis.* 11: 61-67.

621 Cummings, K.J., L.D. Warnick, M. Elton, Y.T. Gröhn, P.L. McDonough, and J.D. Siler. 2010.
622 The effect of clinical outbreaks of salmonellosis on the prevalence of fecal *Salmonella*
623 shedding among dairy cattle in New York. *Foodborne Pathog. Dis.* 7: 815-823.

624 Curiale, M. MPN Calculator, Build 23. <http://i2workout.com/mcuriale/mpn/index.html>.

625 Dodson, K. and J. LeJeune. 2005. *Escherichia coli* O157:H7, *Campylobacter jejuni*, and
626 *Salmonella* prevalence in cull dairy cows marketed in northeastern Ohio. J. Food Prot.
627 68: 927-931.

628 Duffy, G. 2003. Verocytotoxic *Escherichia coli* in animal faeces, manures and slurries. J.
629 Appl. Microbiol. 94(S): 94S-103S.

630 Duse, A., K.P. Waller, U. Emanuelson, H.E. Unnerstad, Y. Persson, and B. Bengtsson. 2014.
631 Risk factors for antimicrobial resistance in fecal *Escherichia coli* from preweaned dairy
632 calves. J. Dairy Sci. 98: 1-17.

633 DWR, 2004. San Joaquin Valley Groundwater Basin, Modesto Subbasin, San Joaquin River
634 Hydrologic Region, California's Groundwater, Bulletin 118. California Department of
635 Water Resources, Online Description.

636 DWR, 2012. Lines of Equal Depth to Water in Wells, Unconfined Aquifer, San Joaquin Valley,
637 Spring 2010 Map. State of California, Department of Water Resources, San Joaquin
638 District.

639 Edberg, S.C., H. LeClerc, and J. Robertson. 1997. Natural protection of spring and well drinking
640 water against surface microbial contamination. II. Indicators and monitoring parameters
641 for parasites. Crit. Rev. Microbiol. 23: 179-206.

642 Edrington, T.S., W.E. Fox, T.R. Callaway, R.C. Anderson, D.W. Hoffman, and D. J. Nisbet.
643 2009. Pathogen prevalence and influence of composted dairy manure application on
644 antimicrobial resistance profiles of commensal soil bacteria. Foodborne Pathog. Dis. 6:
645 217-224.

646 Embrey, S.S., and D.L. Runkle. 2006. Microbial quality of the nation's ground-water resources.
647 United States Geological Survey, National Water-Quality Assessment Program Principal

648 Aquifers. 1993-2004, Scientific Investigations Report 2006-5290, U.S. Department of the
649 Interior, 2006. <http://pubs.water.usgs.gov/sir20065290>.

650 Esiobu, N., L. Armenta, and J. Ike. 2002. Antibiotic resistance in soil and water environments.
651 Int. J. Environ. Health Res. 12: 133-144.

652 Faunt, C.C., K. Belitz, and R.T. Hanson. 2009. Groundwater availability of the Central Valley
653 aquifer, California. Professional Paper 1766, United States Geological Survey.

654 Ferguson, A.S., A.C. Layton, B.J. Mailloux, P.J. Culligan, D.E. Williams, A.E. Smartt, G.S.
655 Sayler, J. Feighery, L.D. McKay, P.S. Knappett, E. Alexandrova, T. Arbit, M. Emch, V.
656 Escamilla, K.M. Ahmed, M.J. Alam, P.K. Streatfield, M. Yunus, and A. van Geen. 2012.
657 Comparison of Fecal Indicators with Pathogenic Bacteria and Rotavirus in Groundwater.
658 Sci. Total Environ. 431: 314-322.

659 Fermér, C.H., and E.O. Engvall. 1999. Specific PCR identification and differentiation of the
660 thermophilic campylobacters, *Campylobacter jejuni*, *C. coli*, *C. lari*, and *C. upsaliensis*.
661 J. Clin. Microbiol. 37: 3370-3373.

662 Fessler, A.T., R.G. Olde Riekerink, A. Rothkamp, K. Kadlec, O.C. Sampimon, T.J. Lam, and S.
663 Schwarz. 2012. Characterization of methicillin-resistant *Staphylococcus aureus* CC398
664 obtained from humans and animals on dairy farms. Vet. Microbiol. 160: 77-84.

665 Foppen, J.W.A., and J.F. Schijven. 2006. Evaluation of data from the literature on the transport
666 and survival of *Escherichia coli* and thermotolerant coliforms in aquifers under saturated
667 conditions. Water Res. 40: 401-426. <http://dx.doi.org/10.1016/j.watres.2005.11.018>

668 Friesema, I., J. Van De Kassteele, C. De Jager, A. Heuvelink, and W. Van Pelt. 2011.
669 Geographical association between livestock density and human Shiga toxin-producing
670 *Escherichia coli* O157:H7 infections. Epidemiol. Infect. 139: 1081-1087.

671 Gibbons, J.F., F. Boland, J.F. Buckley, F. Butler, J. Egan, S. Fanning, B.K. Markey, and F.C.
672 Leonard. 2014. Patterns of antimicrobial resistance in pathogenic *Escherichia coli*
673 isolates from cases of calf enteritis during the spring-calving season. *Vet. Microbiol.* 170:
674 73-80.

675 Gibson, K.E., and K.J. Schwab. 2011a. Detection of Bacterial Indicators and Human and Bovine
676 Enteric Viruses in 100 L Surface and Ground Water Samples Potentially Impacted by
677 Animal and Human Wastes in Lower Yakima Valley, Washington. *Appl. Environ.*
678 *Microbiol.* 77: 355-362.

679 Gibson, K.E., and K.J. Schwab. 2011b. Tangential Flow Ultrafiltration with Integrated Inhibition
680 Detection for the Recovery of Surrogates and Human Pathogens from Large-Volume
681 Source and Finished Drinking Water. *Appl. Environ. Microbiol.* 77: 385-391.

682 Goss, M., D. Barry, and D. Rudolph. 1998. Contamination in Ontario farmstead domestic wells
683 and its association with agriculture: 1. Results from drinking water wells. *J. Contam.*
684 *Hydrol.* 32: 267-293.

685 Goss, M.J., K.S. Rollins, K. McEwan, J.R. Shaw, and H. Lammers-Helps. 2002. The
686 management of manure in Ontario with respect to water quality (Commissioned Paper
687 No. 6). In: *The Walkerton Inquiry*. Queen's Printer for Ontario. Ontario Ministry of the
688 Attorney General, Toronto, Ont.

689 Gronberg, J.M., N.M. Dubrovsky, C.R. Kratzer, J.L. Domagalski, L.R. Brown, and K.R. Burow.
690 1998. Environmental settings of the San Joaquin-Tulare Basins, California. *Water-*
691 *resources Investigations Report 97-4205*. U.S. Geological Survey.

692 Harter, T., H. Davis, M. C. Mathews, R. D. Meyer. 2002. Shallow groundwater quality on dairy
693 farms with irrigated forage crops, *Journal of Contaminant Hydrology* 55 (3-4), pp. 287-
694 315.

695 Harter T., N. Watanabe, X. Li, E.R. Atwill, and W. Samuels. 2014. Microbial groundwater
696 sampling protocol for fecal-rich environments. *Ground Water*. 52(Suppl 1): 126-136.

697 Hill, V.R., L. A. L. Polaczyk, D. Hahn, J. Narayanan, T. L. Cromeans, J. M. Roberts, and J. E.
698 Amburgey. 2005. Development of a rapid method for simultaneous recovery of diverse
699 microbes in drinking water by ultrafiltration with sodium polyphosphate and surfactants.
700 *Appl. Environ. Microbiol.* 71: 6878-6884.

701 Hoar, B.R., E.R. Atwill, C. Elmi, W.W. Utterback, and A. J. Edmondson. 1999. Comparison of
702 fecal samples collected per rectum and off the ground for estimation of environmental
703 contamination attributable to beef cattle. *Am. J. Vet. Res.* 60: 1352-1356.

704 Holmes, M.A., and R.N. Zadoks. 2011. Methicillin resistant *S. aureus* in human and bovine
705 mastitis. *J. Mammary Gland Biol. Neoplasia.* 16: 373-382.

706 Horn, J. and T. Harter. 2009. Domestic well capture zone and influence of the gravel pack
707 length. *Ground Water* 47(2):277-286.

708 Joy, D.M., H. Lee, C. M. Reaume, H.R. Whiteley, and S. Zelin. 1998. Microbial contamination
709 of subsurface tile drainage water from field applications of liquid manure. *Can. Agric.*
710 *Eng.* 40: 153-160.

711 Kirkhorn, S.R. 2002. Community and environmental health effects of concentrated animal
712 feeding operations. *Minn. Med.* 85: 38-43.

713 Knappett, P.S.K., L.D. McKay, A. Layton, D.E. Williams, M.J. Alam, M.R. Huq, J. Mey, J.E.
714 Feighery, P.J. Culligan, B.J. Mailloux, J. Zhuang, V. Escamilla, M. Emch, E. Perfect,

715 G.S. Sayler, K.M. Ahmed, and A. van Geen. 2012. Implications of fecal bacteria input
716 from latrine-polluted ponds for wells in sandy aquifers. *Environ. Sci. Technol.* 46: 1361-
717 1370.

718 Koike, S., I.G. Krapac, H.D. Oliver, A.C. Yannarell, J.C. Chee-Sanford, R.I. Aminov, and R.I.
719 Mackie. 2007. Monitoring and source tracking of tetracycline resistance genes in lagoons
720 and groundwater adjacent to swine production facilities over a 3-year period. *Appl.*
721 *Environ. Microbiol.* 73: 4813-4823.

722 Kummerer, K. 2003. Significance of antibiotics in the environment. *J. Antimicrob. Chemother.*
723 52: 5-7.

724 Lewis, D.J., E.R. Atwill, M.S. Lennox, L. Hou, B. Karle, and K.W. Tate. 2005. Linking on-farm
725 dairy management practices to storm-flow fecal coliform loading for California coastal
726 watersheds. *Environ. Monit. Assess.* 107: 407-425.

727 Li X., N. Watanabe, C. Xiao, T. Harter, B. McCowan, Y. Liu, and E.R. Atwill. 2014. Antibiotic-
728 resistant *E. coli* in surface water and groundwater in dairy operations in Northern
729 California. *Environ. Monit. Assess.* 186: 1253-1260.

730 Lindeman, C.J., E. Portis, L. Johansen, L.M. Mullins, and G.A. Stoltman. 2013. Susceptibility to
731 antimicrobial agents among bovine mastitis pathogens isolated from North American
732 dairy cattle, 2002-2010. *J. Vet. Diagn. Invest.* 25: 581-591.

733 Lockhart, K.M., A.M. King, and T. Harter. 2013. Identifying sources of groundwater nitrate
734 contamination in a large alluvial groundwater basin with highly diversified intensive
735 agricultural production. *J. Contam. Hydrol.* 151: 140-154.

736 Mackie, R.I., S. Koike, I. Krapac, J. Chee-Sanford, S. Maxwell, and R.I. Aminov. 2006.
737 Tetracycline residues and tetracycline resistance genes in groundwater impacted by swine
738 production facilities. *Anim. Biotechnol.* 17: 157-176.

739 McKeon, D.M., J. P. Calabrese, and G.K. Bissonnette. 1995. Antibiotic resistant gram-negative
740 bacteria in rural groundwater supplies. *Water Research.* 29: 1902-1908.

741 Moriarty, E.M., M.L. Macheknzie, N. Karki, and L.W. Sinton. 2011. Survival of *Escherichia*
742 *coli*, *Enterococci*, and *Campylobacter* spp. in Sheep Feces on Pastures. *Appl Environ*
743 *Microbiol.* doi:10.1128/AEM.01329-10.

744 Murinda, S.E., L.T. Nguyen, H.M. Nam, R.A. Almeida, S.J. Headrick, and S.P. Oliver. 2004.
745 Detection of sorbitol-negative and sorbitol-positive Shiga toxin-producing *Escherichia*
746 *coli*, *Listeria monocytogenes*, *Campylobacter jejuni*, and *Salmonella* spp. in dairy farm
747 environmental samples. *Foodborne Pathog. Dis.* 1: 97-104.

748 Nicholson, F. A., S. J. Groves, B. J. Chambers, 2005. Pathogen survival during livestock manure
749 storage and following land application. *Bioresource Technol* 96:135-143.

750 Page, R.W, 1986. Geology of the fresh ground-water basin of the Central Valley, California,
751 with texture maps and sections: U.S. Geological Survey Professional Paper 1401-C, 54 p.

752 Pang, L. 2009. Microbial removal rates in subsurface media estimated from published studies of
753 field experiments and large intact soil cores. *J. Environ. Qual.* 38: 1531-1559.

754 Park, Y., E.R. Atwill, L. Hou, A.I. Packman and T. Harter. 2012. Deposition of *Cryptosporidium*
755 *parvum* oocysts in porous media: a synthesis of attachment efficiencies measured under
756 varying environmental conditions. *Environ. Sci. Technol.* 46: 9491-9500.

757 Paton, A.W., and J.C. Paton. 2003. Direct detection and characterization of Shiga toxigenic
758 *Escherichia coli* by multiplex PCR for stx1, stx2, eae, ehxA, and saa. J. Clin. Microbiol.
759 40: 271-274.

760 Pitkänen, T., P. Karinen, I. Miettinen, H. Lettojärvi, A. Heikkilä, R. Maunula, V. Aula, H.
761 Kuronen, A. Vepsäläinen, L. Nousiainen, S. Pelkonen, and H. Heinonen-Tanski. 2011.
762 Microbial contamination of groundwater at small community water supplies in Finland.
763 *Ambio*. 40: 377-390.

764 Pradhan, A. K., J.S. Van Kessel, J. S. Karns, D.R. Wolfgang, E. Hovingh, K.A. Nelen, J.M.
765 Smith, R.H. Whitlock, T. Fyock, S. Ladely, P.J. Fedorka-Cray, and Y.H. Schukken.
766 2009. Dynamics of endemic infectious diseases of animal and human importance on three
767 dairy herds in the northeastern United States. *Journal of Dairy Science*, 92:1811-1825.

768 Purdy, C.W., D.C. Straus, D.B. Parker, B.P. Williams, and R.N. Clark. 2001. Water quality in
769 cattle feedyard playas in winter and summer. *Am. J. Vet. Res.* 62: 1402-1407.

770 Ravva, S.V. and C.Z. Sarreal. 2014. Survival of *Salmonella enterica* in aerated and nonaerated
771 wastewaters from dairy lagoons. *Int. J. Environ. Res. Public Health*. 11: 11249-11260.

772 Rudolph, D.L., D.A.J. Barry, and M.J. Goss. 1998. Contamination in Ontario farmstead domestic
773 wells and its association with agriculture: 2. Results from multilevel monitoring well
774 installations. *J. Contam. Hydrol.* 32: 295-311.

775 Saini, V., J.T. McClure, D.T. Scholl, T.J. DeVries, and H.W. Barkema. 2013. Herd-level
776 relationship between antimicrobial use and presence or absence of antimicrobial
777 resistance in gram-negative bovine mastitis pathogens on Canadian dairy farms. *J. Dairy*
778 *Sci.* 96: 4965-4976.

779 Sapkota, A.R., F.C. Curriero, K.E. Gibson, and K.J. Schwab. 2007. Antibiotic-resistant
780 enterococci and fecal indicators in surface water and groundwater impacted by a
781 concentrated swine feeding operation. *Environ. Health Perspect.* 115: 1040-1045.

782 Searcy, K.E., A.I. Packman, E.R. Atwill and T. Harter. 2005. Association of *Cryptosporidium*
783 *parvum* with suspended particles: impact on oocyst sedimentation. *Appl. Environ.*
784 *Microbiol.* 71: 1072-1078.

785 Shere, J.A., K.J. Bartlett, and C.W. Kaspar. 1998. Longitudinal study of *Escherichia coli*
786 O157:H7 dissemination on four dairy farms in Wisconsin. *Appl. Environ. Microbiol.* 64:
787 1390-1399.

788 Sinton, L.W., R.R. Braithwaite, C.H. Hall, and M.L. Mackenzie. 2007. Survival of Indicator and
789 Pathogenic Bacteria in Bovine Feces on Pasture. *Appl Environ Microbiol.* 73(24): 7917-
790 25.

791 Straley, B.A., S.C. Donaldson, N.V. Hedge, A.A. Sawant, V. Srinivasan, S.P. Oliver, and B.M.
792 Jayarao. 2006. Public health significance of antimicrobial-resistant gram-negative
793 bacteria in raw bulk tank milk. *Foodborne Pathog. Dis.* 3: 222-233.

794 Toth, J.D., H.W. Aceto, S.C. Rankin, and Z. Dou. 2011. Survival characteristics of *Salmonella*
795 *enterica* serovar Newport in the dairy farm environment. *J. Dairy Sci.* 94: 5238-5246.

796 Toth, J.D., H.W. Aceto, S.C. Rankin, and Z. Dou. 2013. Short communication: Survey of
797 animal-borne pathogens in the farm environment of 13 dairy operations. *J. Dairy Sci.* 96:
798 5756-5761.

799 Unc, A., M. J. Goss, S. Cook, X. Li, E. R. Atwill, and T. Harter. 2012. Analysis of matrix effects
800 critical to microbial transport in organic waste-affected soils across laboratory and field
801 scales. *Water Resour. Res.* 48, W00L12, 17p., doi:10.1029/2011WR010775.

802 Unc, A., and M. Goss. 2014. Transport of bacteria from manure and protection of water
803 resources. Appl. Soil Ecology. 25: 1-18.

804 United States Environmental Protection Agency, 2012. Region 9 strategic plan, 2011-2014.
805 Technical Report.

806 United States Environmental Protection Agency, 1993. Method 2007, Trace Elements in Water,
807 Solids, and Biosolids by Inductively Coupled Plasma-Atomic Emission Spectrometry,
808 [http://water.epa.gov/scitech/methods/cwa/bioindicators/upload/2007_07_10_methods_me](http://water.epa.gov/scitech/methods/cwa/bioindicators/upload/2007_07_10_methods_method_2007.pdf)
809 [thod_2007.pdf](http://water.epa.gov/scitech/methods/cwa/bioindicators/upload/2007_07_10_methods_method_2007.pdf) (downloaded 6/8/2015). U.S. EPA Method 300.0, Determination of
810 Inorganic Anions by Ion Chromatography, Revision 2.1.

811 USDA APHIS Antibiotic Use on U.S. Dairy Operations, 2002 and 2007. Retrieved from
812 [http://www.aphis.usda.gov/animal_health/nahms/dairy/downloads/dairy07/Dairy07_is_](http://www.aphis.usda.gov/animal_health/nahms/dairy/downloads/dairy07/Dairy07_is_AntibioticUse.pdf)
813 [AntibioticUse.pdf](http://www.aphis.usda.gov/animal_health/nahms/dairy/downloads/dairy07/Dairy07_is_AntibioticUse.pdf).

814 Van der Schans, M. L., T. Harter, A. Leijnse, M. C. Mathews, R. D. Meyer. 2009. Characterizing
815 sources of nitrate leaching from an irrigated dairy farm in Merced County, California, J.
816 of Contam. Hydrology 110:9-21.

817 Ward, M.J., C.L. Gibbons, P.R. McAdam, B.A. van Bunnik, E.K. Girvan, G.F. Edwards, J.R.
818 Fitzgerald, and M.E. Woolhouse. 2014. Time-scaled evolutionary analysis of the
819 transmission and antibiotic resistance dynamics of *Staphylococcus aureus* CC398. Appl.
820 Environ. Microbiol. 80: 7275-7282.

821 Watanabe, N., B.A. Bergamaschi, K.A. Loftin, M.T. Meyer, and T. Harter. 2010. Use and
822 environmental occurrence of antibiotics in freestall dairy farms with manured forage
823 fields. Environ. Sci. Technol. 44: 6591-6600.

824 Watson, E., S. Jeckel, L. Snow, R. Stubbs, C. Teale, H. Wearing, R. Horton, M. Toszeghy, O.
825 Tearne, J. Ellis-Iversen, and N. Coldham. 2012. Epidemiology of extended spectrum
826 beta-lactamase *E. coli* (CTX-M-15) on a commercial dairy farm. *Vet. Microbiol.* 154:
827 339-346.

828 Watts, J.L., T.R. Shryock, M. Apley, D.J. Bade, S.D. Brown, J.T. Gray, H. Heine, R.P. Hunter,
829 D.J. Mevius, M.G. Papich, P. Silley, and G.E. Zurenko. 2008. Performance Standards for
830 Antimicrobial Disk and Dilutions Susceptibility Tests for Bacteria Isolated From
831 Animals; Approved Standard-Third Edition. Clinical and Laboratory Standards Institute.
832 M31-A3 (ISBN 1-56238-659-X).

833 Wichmann, F., N. Udikovic-Kolic, S. Andrew, and J. Handelsman. 2014. Diverse antibiotic
834 resistance genes in dairy cow manure. *Mbio.* 5(2): e01017-13.

835 Wieler, L.H. 2014. "One Health" – Linking human, animal and environmental health. *Int. J.*
836 *Med. Microbiol.* 304: 775-776.

837 Wilhelm, B., A. Rajić, L. Waddell, S. Parker, J. Harris, K.C. Roberts, R. Kydd, J. Greig, and A.
838 Baynton. 2009. Prevalence of zoonotic or potentially zoonotic bacteria, antimicrobial
839 resistance, and somatic cell counts in organic dairy production: current knowledge and
840 research gaps. *Foodborne Pathog. Dis.* 6: 525-539.

841 Witte, W. 1998. Medical consequences of antibiotic use in agriculture. *Science.* 279: 996-997.

842 Won, G., A. Gill, and J. LeJeune. 2013. Microbial quality and bacteria pathogens in private wells
843 used for drinking water in northeastern Ohio. *J. Water Health.* 11: 555-562.

844

845

846

Table 1. Survey of generic *E. coli* in groundwater from CAFOs wells located in the Central Valley, California (2008-2009).

Dairy ID	January 2008		April 2008		September 2008		March-April 2009	
	% (positive /sampled wells)	Mean concen. [†] (±SD)	% (positive /sampled wells)	Mean concen. [†] (±SD)	% (positive /sampled wells)	Mean concen. [†] (±SD)	% (positive /sampled wells)	Mean concen. [†] (±SD)
Monitoring wells (N=190)								
36-04	0 (0/3)	NA	0 (0/3)	NA	0 (0/4)	NA	0 (0/4)	NA
36-11	0 (0/3)	NA	33.3 (1/3)	0.02	ND	NA	0 (0/1)	NA
36-15	0 (0/10)	NA	20.0 (2/10)	0.03 (0.007)	21.4 (3/14)	0.02 (0.006)	14.3 (2/14)	0.01(0)
36-19	9.1 (1/11)	5.45	36.4 (4/11)	0.04 (0.06)	23.1 (3/13)	0.01 (0.005)	8.3 (1/12)	0.05
36-24	0 (0/2)	NA	ND	NA	ND	NA	ND	NA
36-27	0 (0/3)	NA	66.7 (2/3)	0.70 (0.98)	0 (0/3)	NA	0 (0/3)	NA
37-39	60.0 (3/5)	35.01 (30.30)	33.3 (2/6)	0.48 (0.65)	100 (6/6)	1.58 (3.64)	87.5 (7/8)	3.24 (8.40)
37-42	33.3 (3/9)	2.24 (3.78)	12.5 (1/8)	0.04	11.1 (1/9)	4.30	44.4 (4/9)	0.75 (1.47)
Overall	15.2 (7/46)	16.74 (24.58)	27.3 (12/44)	0.22 (0.45)	26.5 (13/49)	1.06 (2.66)	27.5 (14/51)	1.84 (5.93)
Domestic wells [‡] (N=24)								
36-04	ND	NA	ND	NA	0 (0/1)	NA	0 (0/1)	NA
36-15	ND	NA	ND	NA	0 (0/1)	NA	0 (0/1)	NA
36-19	0 (0/1)	NA						
36-27	0 (0/1)	NA						
37-39	33.3 (1/3)	0.01	0 (0/3)	NA	0 (0/3)	NA	0 (0/3)	NA
Overall	20.0 (1/5)	0.01	0 (0/5)	NA	0 (0/7)	NA	0 (0/7)	NA

[†] Concentrations were expressed as CFU/100-ml; [‡] No domestic wells were sampled on dairies 36-11, 36-24 and 37-42; NA=Not applicable; ND=Not done (due to the well was not accessible or no water available).

Table 2. Survey of *Enterococcus* spp. in groundwater from CAFO wells in the Central Valley, California (2008-2009).

Dairy ID	January 2008		April 2008		September 2008		March-April 2009	
	% (positive /sampled wells)	Mean concen. [†] (±SD)	% (positive /sampled wells)	Mean concen. [†] (±SD)	% (positive /sampled wells)	Mean concen. [†] (±SD)	% (positive /sampled wells)	Mean concen. [†] (±SD)
Monitoring wells (N=190)								
36-04	100 (3/3)	6.71 (6.09)	100 (3/3)	0.37 (0.34)	100 (4/4)	4.75 (6.05)	100 (4/4)	3.58 (4.48)
36-11	100 (3/3)	7.08 (6.36)	100 (3/3)	5.11 (3.62)	ND	NA	100 (1/1)	13.60
36-15	100 (10/10)	2.83 (3.67)	100 (10/10)	5.52 (8.65)	85.7 (12/14)	19.49 (48.13)	100 (14/14)	11.99 (11.34)
36-19	100 (11/11)	8.87 (7.08)	100 (11/11)	3.22 (3.73)	100 (13/13)	4.87 (5.72)	100 (12/12)	5.11 (4.77)
36-24	0 (0/2)	NA	ND	NA	ND	NA	ND	NA
36-27	100 (3/3)	11.82 (11.20)	66.7 (2/3)	3.14 (4.33)	100 (3/3)	26.56 (21.05)	100 (3/3)	5.57 (4.12)
37-39	100 (5/5)	43.19 (56.28)	100 (6/6)	30.14 (43.93)	100 (6/6)	22.15 (10.60)	100 (8/8)	3822.88 (7181.30)
37-42	100 (9/9)	15.21 (21.06)	87.5 (7/8)	4.58 (6.02)	88.9 (8/9)	36.92 (90.24)	100 (9/9)	263.59 (726.44)
Overall	100 (46/46)	12.63 (23.27)	95.5 (42/44)	7.29 (17.21)	93.9 (46/49)	17.33 (45.38)	100 (51/51)	651.55 (3036.82)
Domestic wells [‡] (N=24)								
36-04	ND	NA	ND	NA	100 (1/1)	0.38	100 (1/1)	0.30
36-15	ND	NA	ND	NA	100 (1/1)	0.35	100 (1/1)	0.60
36-19	100 (1/1)	0.80	100 (1/1)	3.66	100 (1/1)	0.28	100 (1/1)	1.30
36-27	100 (1/1)	4.59	100 (1/1)	0.27	100 (1/1)	3.12	100 (1/1)	0.40
37-39	66.7 (2/3)	2.39 (0.26)	66.7 (2/3)	0.10 (0.05)	66.7 (2/3)	0.42	100 (3/3)	7.0 (6.0)
Overall	80.0 (4/5)	2.54 (1.56)	80.0 (4/5)	1.03 (1.78)	85.7 (6/7)	0.91 (1.24)	100 (7/7)	3.37 (4.86)

[†] Concentrations were expressed as CFU/100-ml; [‡] No domestic wells were sampled on dairies 36-11, 36-24 and 37-42; NA=Not applicable; ND=Not done (due to the well was not accessible or no water available).

Table 3. Factors associated to the concentrations of generic *E. coli* and *Enterococcus* spp. in groundwater in CAFOs in the Central Valley, California (2008-2009).

Bacteria	Factor	Coef.	Std. Err.	P value	95% CI
<i>Generic E. coli</i>	Land use				
	Corral [†]	0	--	--	--
	Field	-2.824	0.331	0.000	(-3.473, -2.175)
	Lagoon	-2.008	0.484	0.000	(-2.958, -1.058)
	Upgradient	-6.098	2.887	0.035	(-11.758, -0.438)
	Water table depth				
	Deep [†]	0	--	--	--
	Shallow	3.825	0.381	0.000	(3.077, 4.572)
	Season [‡]				
	Spring [†]	0	--	--	--
	Fall	0.436	0.360	0.226	(-0.269, 1.141)
Winter	2.734	0.255	0.000	(2.233, 3.235)	
<i>Enterococcus</i>	Well type				
	Monitoring well [†]	0	--	--	--
	Domestic well	-3.536	0.155	0.000	(-3.839, -3.233)
	Land use				
	Corral [†]	0	--	--	--
	Field	-1.239	0.021	0.000	(-1.280, -1.197)
	Lagoon	1.836	0.011	0.000	(1.812, 1.859)
	Upgradient	-3.322	0.090	0.000	(-3.500, -3.145)
	Water table depth				
	Deep [†]	0	--	--	--
	Shallow	4.109	0.032	0.000	(4.046, 4.172)
	Season [‡]				
	Spring [†]	0	--	--	--
Fall	-2.828	0.036	0.000	(-2.899, -2.757)	
Winter	-5.874	0.056	0.000	(-5.984, -5.764)	

[†] Referent category for categorical variable.

[‡] Spring: April 2008 and March/April 2009; Fall: September 2008; Winter: January 2008.

Table 4. Survey of indicator and pathogenic bacteria in groundwater from domestic wells with and without likely dairy influence, in the Central Valley, California (2010 - 2011).

Well types	Bacteria	No. (%) of positive wells	Concentration (CFU/100ml) Mean (\pm SD)
Near-Dairy Wells [†] (N=132)			
within dairy facility (N=12)	Generic <i>E. coli</i>	1 (8.3)	0.72
	<i>Enterococcus</i> spp.	3 (25.0)	0.13 (0.15)
	<i>Campylobacter</i> spp.	0 (0)	NA
	<i>Salmonella</i> spp.	0 (0)	NA
	<i>E. coli</i> O157:H7	0 (0)	NA
outside of dairy facilities (N=120)	Generic <i>E. coli</i>	5 (4.2)	0.26 (0.39)
	<i>Enterococcus</i> spp.	7 (5.8)	16.93 (43.33)
	<i>Campylobacter</i> spp.	0 (0)	NA
	<i>Salmonella</i> spp.	0 (0)	NA
	<i>E. coli</i> O157:H7	0 (0)	NA
Non-Dairy Wells (without dairy influence) [†] (N=68)			
	Generic <i>E. coli</i>	4 (5.9)	1.93 (3.33)
	<i>Enterococcus</i> spp.	7 (10.3)	15.03 (37.15)
	<i>Campylobacter</i> spp.	0 (0)	NA
	<i>Salmonella</i> spp.	0 (0)	NA
	<i>E. coli</i> O157:H7	0 (0)	NA

[†] Wells are defined as “near-dairy” if the distance between a well and a dairy lagoon or corral is 2.4km or less; wells are defined as “non-dairy” if the distance between a well and a dairy lagoon or corral is greater than 2.4km; NA=Not applicable.

853

854

855

Table 5. Factors associated to the concentrations of *Enterococcus* spp. in groundwater from domestic wells in the Central Valley, California (2010-2011).

Factor	Coef.	Std. Err.	P value	95% CI
Well designation				
Non-Dairy well [†]	0	--	--	--
Near-Dairy well	1.900	0.761	0.013	(0.408, 3.393)
Potassium	-29.955	11.284	0.008	(-52.071, -7.839)
Season				
Spring [†]	0	--	--	--
Summer	-2.120	4.935	0.667	(-11.793, 7.552)
Fall	1.922	0.853	0.024	(0.251, 3.593)
Winter	0.967	1.043	0.354	(-1.077, 3.010)

[†] Referent category for categorical variable

857

858

859

860

861

862

863

864

865

866

867

868

869

Table 6. Number of antibiotic resistant isolates of generic *E. coli* and *Enterococcus* spp. isolated from groundwater

Bacteria	Campaign	Type of wells	No. isolates tested	No. isolates [†] resistant to ≥ 1 antibiotics	No. isolates [†] resistant to ≥ 3 antibiotics	
<i>E. coli</i>	On-dairy	Monitoring wells	2	2	2	
		Domestic wells	0	NA	NA	
	2010-2011	Near-dairy wells [‡]	5	5	3	
		Non-dairy wells [‡]	4	4	2	
	Total			11	11	7
	<i>Enterococcus</i>	On-dairy	Monitoring wells	18	18	16
Domestic wells			4	4	2	
2010-2011		Near-dairy wells [‡]	8	8	8	
		Non-dairy wells [‡]	6	6	5	
Total			36	36	31	

[†]Each tested isolate was from different wells; [‡] Near-dairy” wells are located at a distance not exceeding 2.4 km from the nearest dairy lagoon or corral; wells are otherwise defined as “non-dairy”; NA=Not applicable.

871

872

873

874

875

876

877

878

879

880

Figure Captions

881

882

883 **Figure 1.** Study area in California, USA focusing on four counties overlying the Central Valley
884 aquifer (dark grey): Stanislaus County (S), Merced County (M), Kings County (K),
885 and Tulare County (T). The southern and central Central Valley has a high
886 concentration of dairy CAFOs and manure-treated irrigated crops (black). The study
887 was conducted on eight dairies (red) and also included a regional survey of 200
888 domestic wells (white).

889 **Figure 2.** Concentration of generic *E. coli* (A), *Enterococcus* (B), *Campylobacter* spp. (C), and
890 *Salmonella* spp. (D) in surface solids of various management units on two San Joaquin
891 Valley dairies in the Stanislaus and Merced County portion of our study area. These
892 schematics represent scaled version of the dairy management units. Concentrations of
893 these bacteria in flush water, calf hutches flush water, and in storage lagoons is shown
894 in the panel (E).

895 **Figure 3.** Frequencies of detection of generic *E. coli* and *Enterococcus* spp. in groundwater from
896 monitoring wells proximity to different types of land on CAFOs in the Central Valley,
897 California (2008-2009).